Pacific Conservation Biology, 2019, **25**, 331–341 https://doi.org/10.1071/PC18079

Overlooked and undervalued: the neglected role of fauna and a global bias in ecological restoration assessments

Sophie L. Cross A,F, Sean Tomlinson A,B, Michael D. Craig C,D, Kingsley W. Dixon and Philip W. Bateman E

Abstract. Globally increasing rates of mine site discontinuations are resulting in the need for immediate implementation of effective conservation and management strategies. Surveying vegetation structure is a common method of assessing restoration success; however, responses of fauna to mine site restoration remain largely overlooked and understudied despite their importance within ecosystems as ecological engineers, pollinators, and restoration facilitators. Here we review the current state of the use of fauna in assessments of mine site restoration success globally, and address biases or shortcomings that indicate the assessment approach may undershoot closure and restoration success. We identified just 101 peer-reviewed publications or book chapters over a 49-year period that assess responses of fauna to mine site restoration globally. Most studies originate in Australia, with an emphasis on just one company. Assessments favour general species diversity and richness, with a particular focus on invertebrate responses to mine site restoration. Noteworthy issues included biases towards origin of study, study type, and target taxa. Further searches of the grey literature relating to fauna monitoring in mine site restoration, which was far more difficult to access, yielded six monitoring/guidance documents, three conference proceedings, two book chapters without empirical data, and a bulletin. As with peer-reviewed publications, grey literature focussed on invertebrate responses to restoration, or mentioned fauna only at the most basic level. We emphasise the need for global re-evaluation of regulatory standards to address these major limitations in assessing the capacity of the mining industry to comprehensively and representatively restore faunal communities after mining.

Additional keywords: mine, rehabilitation

Received 15 October 2018, accepted 16 December 2018, published online 25 January 2019

Introduction

Habitat destruction and fragmentation are primary drivers of biodiversity loss and extinctions worldwide, and the effects of these are being increasingly exacerbated through human activities such as mining, agriculture, forestry and urbanisation (Fahrig 1997; Lande 1998; Tilman *et al.* 2001; Cristescu *et al.* 2012). While the physical environmental footprint of mining operations is <1% of terrestrial landscape areas, and relatively concentrated in comparison to other industries, e.g. agriculture and urbanisation, which account for 70% and 3% of global land disturbances, respectively (Hodges 1995; Bridge 2004; McKinney 2006), mining often has a substantial local, and often regional, environmental impact (Salomons 1995; Rybicka

1996). Activities from mining can fundamentally alter relatively intact and undisturbed habitats into inhospitable land matrices, and can create serious environmental pollution issues such as tailings leakage, dust, and hydrological change (Salomons 1995; Bian *et al.* 2009). Though mining activities impact a small terrestrial footprint, 75% of active sites are situated on land considered to be of high conservation value (Miranda *et al.* 2003; Bridge 2004). Hence, although environmental impacts of mineral extraction may be restricted in spatial extent, they are intensely disruptive to ecosystems that are often uncommon and fragile. The resultant alteration and degradation from mining activities present some of the most difficult landscapes to restore. As such, lessons learned from the restoration of mine

^AARC Centre for Mine Site Restoration, School of Molecular and Life Sciences, Curtin University, Kent Street, Bentley, WA 6102, Australia.

^BKings Park Science, Department of Biodiversity Conservation and Attractions, Kattij Close, Kings Park, WA 6005, Australia.

^CSchool of Biological Sciences, University of Western Australia, Crawley, WA 6009, Australia.

^DSchool of Veterinary and Life Sciences, Murdoch University, Murdoch, WA 6150, Australia.

^ESchool of Molecular and Life Sciences, Curtin University, Kent Street, Bentley, WA 6102, Australia.

^FCorresponding author: sophie.cross@postgrad.curtin.edu.au

sites may be transferrable to land restoration practices in other areas of high conservation value that have suffered other forms of degrading processes.

Many different environmental components (e.g. soil, plants, microorganisms, and fauna) require study in assessments of ecosystem health and functionality (Duffy 2003); yet restoration monitoring is typically restricted to plant communities and vegetation structure, which remain a key priority in assessing postmining restoration success (Ruiz-Jaen and Mitchell Aide 2005; Koch et al. 2010). Majer (1989) highlights this issue; however, the disparity between fauna and plant studies remains a key issue. This is despite fauna being essential to restoration success, and playing critical roles in the provision of numerous essential ecosystem functions, such as seed dispersal, pollination, nutrient cycling, and soil formation (Majer 1989; Lavelle et al. 2006; Mace et al. 2012). Importantly, fauna, due to their mobility, often rely on spatial scales far greater than plants, and hence are often dependant on habitats and resources that occur both within and outside the restoration patch. However, responses of fauna are often overlooked in favour of standardised vegetation surveys, which typically can be achieved rapidly and follow established principles (Ruiz-Jaen and Mitchell Aide 2005). Fauna are often assumed to return to predisturbance diversity and abundances following the return of vegetation (Block et al. 2001; Cristescu et al. 2012) through what is commonly referred to as the 'Field of Dreams' Hypothesis ('build it and they will come': Palmer et al. 1997). In practice, recovering animal biodiversity and community structure are some of the most difficult components to understand, achieve, and assess following the restoration of degraded sites (Cristescu et al. 2012; Perring et al. 2015).

Faunal responses to mine site restoration require study across a wide range of habitats and climatic regions to maximise biodiversity outcomes. Biases to certain regions or mineral extraction types limit our ability to inform on best practices for restoring ecosystem function by preconditioning our expectations to outcomes that may be unique to some places or disturbance patterns. Surface (e.g. strip mining, open pit, and quarry) and subsurface (underground) mining have varying levels of physical environmental impact (Dudka and Adriano 1997). Underground mining can have significant impacts on subsurface hydrology and soil structure (Altun et al. 2010); however, the above-ground impact (other than infrastructure and tailings or waste rock dumps) of underground mining is of a lower magnitude by comparison to the often very large terrestrial footprints of surface mining (Lin et al. 2005). Hence, conclusions drawn from sites of only one extraction type may not be best suited to inform restoration practices for other mining techniques.

Faunal responses to mine site restoration also require studies across varying climatic regions. Many of the world's 35 global biodiversity hotspots are situated within the tropics (Mittermeier et al. 2011). These regions contain higher proportions of endemic species than areas outside the hotspots (Myers et al. 2000). Endemic species, by virtue of occupying one or few specialised habitats, are likely to be affected more severely by habitat fragmentation and loss than generalist species, increasing the difficulty associated with restoring biodiversity values and potentially ecosystem functioning (Ewers and Didham

2006). Furthermore, while iron ore extraction from ultramafic soils takes place in biodiverse landscapes in, for example, Brazil, New Caledonia and Australia, it seems unlikely that the best practices of ecological restoration developed in Australia, with its unique flora and fauna and ancient, arid landscapes (Hopper and Gioia 2004; Hopper 2009), would translate well to the different tropical ecosystems of an island in the Pacific, or the rainforests of South America to improve restoration practices and biodiversity conservation.

Although a higher focus is being placed on fauna assessments in restoration in recent years (Majer 2009), of the limited studies that assess animal responses to restoration (particularly in relation to mine site restoration), there is a strong emphasis evident towards the use of certain taxa as biological indicators (bioindicators); for example, ants and birds, both of which typically can be easily surveyed with minimal time and financial investments (Majer 1983; Andersen et al. 2003; Nichols and Nichols 2003; Gould and Mackey 2015). The use of bioindicators has remained a favoured method of assessing environmental health, since the introduction of the concept by Hall and Grinnell (1919). While invertebrates are highly important in ecosystems, and can provide essential information in assessments of environmental health (Majer et al. 2007), basing restoration practices on responses of only ants and common bioindicators may under-represent other groups or negatively affect overall ecosystem development. For restoration efforts to be effective for all faunal groups, assessments for restoration success must be derived from a wider range of fauna, and from their role in the ecosystem, rather than ease of survey effort.

Studies assessing faunal responses to restoration typically favour assessments of species richness and abundance, likely due to reliability and ease of implementation. However, species diversity assessments have several limitations, namely that there is a high probability of missing rare, cryptic, migratory, or seasonally active species, and in the potential for species diversity to be altered through the detection of invasive or cosmopolitan species (Hejda et al. 2009; Chiarucci et al. 2011). Fauna that are capable of dispersing large distances may present a false representation of utilisation of restoration areas, as these areas may only be used opportunistically or transiently and incapable of supporting resident fauna communities in the long term. Isolated assessments of species presence or absence, or diversity, may therefore provide relatively little information as to the functional success of restoration. Studies based primarily on presence or absence do not allow for evaluation of resource use and use of wider restoration landscapes, and hence provide an inaccurate assessment of restoration trajectory and success. Integrative ecological and behavioural studies remain an emerging branch of conservation biology, and might provide an increased understanding of what constitutes a return to a fully restored site. Globally, little is known of how human disturbances alter the behaviour and ecology of fauna that persist in disturbed landscapes, such as postmining environments. Ecological and behavioural studies require significant time investment, and often have higher associated risks and costs than more general species diversity assessments, in terms of the ease of data collection. However, studies of ecology are essential, as behavioural characteristics are the most flexible of faunal adaptations to their environment,

and have differing responses to environmental changes (Wolf and Weissing 2012).

This review assesses the current state of knowledge of the use of fauna in assessments of mine site restoration success. While Cristescu et al. (2012) published a review on the use of fauna in assessments of mining restoration success (termed rehabilitation), they primarily assessed the empirical data on faunal recolonisation of mine sites within Australia, whereas we identify and address any potential biases or patterns within literature assessing faunal responses to mine site restoration on a global scale. Specifically, we assess patterns in origin and year of study, targeted taxa, study type (i.e. presence or absence, or species diversity and abundances), and terminology use. We also seek to extend a similar interrogation to the grey literature surrounding faunal monitoring in mine site restoration. Understanding and addressing the current knowledge gaps in mine site restoration literature allows for the identification of areas requiring an increased study focus, and is integral to implementing the 'International Standards for the Practice of Ecological Restoration' (McDonald et al. 2016).

Methods

We compiled a comprehensive database of peer-reviewed literature composed of studies relating to any use of fauna (invertebrate or vertebrate) in assessments of mining restoration success. Studies were not limited to those using the terminology 'restoration', but included those describing attempted return of vegetation (unassisted natural regeneration or otherwise) following cessation of mining. Mining restoration literature encompasses a wide range of terminologies for describing various restoration practices (Kaźmierczak et al. 2017; Cross et al. 2018). For the purposes of this review, we use 'restoration' (adopted terminology in McDonald et al. 2016), which we define as 'the process of assisting the recovery of an ecosystem that has been degraded, damaged, or destroyed' (Clewell et al. 2004; McDonald et al. 2016). Literature assessing faunal responses to mining without reference to any form of restoration were discounted. We used three databases to interrogate the literature: Google Scholar, Web of Science (all databases, 1950-2018), and Scopus (all documents including secondary documents, all years; last searched November 2018). Additional sources were gleaned from bibliographies in the published literature.

Search terms comprised any combination of 'Australasia', 'Africa', 'North America', 'South America', 'Asia', or 'Europe', AND/OR 'animal', 'fauna', 'bird', 'reptile', 'mammal', 'vertebrate', or 'invertebrate' AND 'response', or 'behaviour' AND 'mine', or 'mining' AND 'restoration', 'rehabilitation', 'reclamation', 'recultivation', 'afforestation', or 'regeneration'. Publications were compiled into a database and sorted based on date of publication, country of origin, target taxa, type of mineral mined, terminology used, and key search terms. The literature comprised 101 publications (references used in analyses but not cited in text are summarised in Table S1). As postmining recovery may not be fully represented in the primary literature, we extracted the grey literature from searches and compiled these into a separate database. Grey literature included unpublished data, articles without empirical data, governmental

reports, conference proceedings, and bulletins (summarised in Tables S1, S2 available as Supplementary Material). Analyses were designed to assess the current state of research in assessments of faunal responses to mining restoration, and identify potential knowledge gaps or biases. Although our aim was to interrogate the grey literature in a similar fashion to the peerreviewed work, our analyses were insupportable due to the paucity of accessible or relevant data.

First, we identified the number of studies from each individual mine site, allowing for the detection of any potential overlaps or biases to particular sites and type of mineral mined. We then grouped studies based on country of origin and year of publication. Third, we identified the main terminology (the primary term used if multiple terms were present) to assess whether there was a standardised approach to terminology. Lastly, we investigated correlations between location, date of publication, and type of study, with use of particular taxa and type of mineral operation. We identified the following seven variables: (1) mineral type: coal (including publications listing the term 'lignite'), bauxite, sand, bentonite, gold, iron ore, limestone, tin, uranium, peat, multiple (polymetallic mines, or mines where two or more mineral types were listed), and not stated; (2) taxon group: vertebrate, invertebrate, or both; (3) target clade: Mammalia, Aves, Reptilia, Amphibia, Insecta, Clitellata (a taxon of annelid worm), or multiple targets; (4) main terminology; (5) date of publication; (6) country of origin; and (7) study type: ecology (pollination, density/ biomass, predation), presence/absence, or population abundance of fauna species, and translocations.

Pearson's Chi-square tests were undertaken to compare differences between all categorical variables. All statistical analyses were conducted in the *R* 3.4.4 statistical environment (R Core Team 2016), implemented using RStudio (RStudio Inc., Boston, United States, 2018). The results from literature searches have been visualised in a PRISMA 2009 flow diagram (Fig. S1, Supplementary Material).

Results

Searches of peer-reviewed, published literature yielded a total of 101 publications from 10 different mineral type operations. Grey literature searches yielded just 12 readily accessible documents, eight of which made direct reference to fauna or fauna monitoring in restoration landscapes. Of the published literature, six studies were based at mines extracting multiple minerals, and five studies did not state the mineral type. Studies predominantly focused on bauxite (n = 34), coal (n = 26), and mineral sand mines (n = 19). Two studies each were from limestone, uranium, gold, and peat mines/quarries, and one each from bentonite, iron ore, and tin mines. Many of these minerals are typically extracted through surface mining, with the exception of coal and gold (both surface and subsurface mining), and uranium (subsurface mining). Terminology varied considerably between publications, with a total of seven different terms used: 'rehabilitation', 'restoration', 'regeneration', 'reclamation', 'recultivation', 'revegetation', and 'afforestation'. Of the 101 publications, 73 used a single terminology to describe restoration activity and 28 mixed terms within the same publication. The countries of origin comprised 14 countries (Australia,

United States, Germany, Brazil, Hungary, Spain, South Africa, New Zealand, Czech Republic, United Kingdom, Canada, Colombia, Indonesia, and Italy), two of which are listed in the top five mineral-producing (by metric ton) countries (Fig. S2*a*, *b*, Supplementary Material). Indonesia, Colombia, Brazil, and Australia are listed in the top five megadiverse countries, ranked 1 to 4, respectively (Fig. S2*c*).

Invertebrate responses to mining restoration were assessed in 60 publications; 39 publications assessed vertebrate responses, and two papers assessed both invertebrate and vertebrate responses. Invertebrate studies favoured assessments for insects (90%), and vertebrate studies typically favoured assessments of birds (46%). Studies were significantly more likely to involve assessments for species diversity and abundance (75%, $\chi^2 = 309.5$, P < 0.001) compared with those including ecology (including pollination, density/biomass, and predation studies; 18%), presence, absence or population abundance of individual species (6%), or translocations (1%).

Terminology

'Rehabilitation' was the most commonly used main term (primary terminology used within the publication; n = 47), followed by 'restoration' (n = 21), 'regeneration' (n = 10), 'reclamation' (n = 8), 'recultivation' (n = 7), 'revegetation' (n = 4), and 'afforestation' (n = 3). The main terminology of one study (either 'restoration' or 'reclamation') could not be ascertained with certainty (Table S3, Supplementary Material). Use of terminology appeared to be, in part, associated with publication date. While 'rehabilitation' had been in consistent use across the range of publication dates (1978 to 2017), 'restoration' appeared to be the favoured term within the last decade. Other terminologies do not appear to be in widespread use. European studies had the widest range of terminology (all terminologies apart from 'regeneration': Table 1). The use of 'afforestation' and 'recultivation' were exclusively restricted to European studies, and 'reclamation' was limited primarily to European and North American studies, with one use in an Australasian study.

Origin and date of study

Studies of fauna in mining restoration were significantly more likely to originate within Australasia than any other region $(59\%, \chi^2 = 293.41, P < 0.001)$. While there is a major Australian bias in the literature, 28 of the 60 Australian studies arise from a single organisation: Alcoa of Australia (hereafter Alcoa), which has extensively reported the role of fauna in the restoration of its bauxite operations in the jarrah forests of southwest Australia. These reports account for 82% of studies of bauxite mines globally (n = 28 of 34), and this pattern is the global norm: many studies within mineral categories result from a single mine site. All eight studies within South Africa are from the same locality (Richards Bay), with similar trends among other countries including Germany (n = 2 of 7, Berzdorf lignite mining district, eastern Germany), Czech Republic (n = 3 of 3, north/north-west Bohemia), Hungary (n = 3 of 3, Pécs, southern Hungary), and New Zealand (n = 2 of 2, Wangaloa coal mine, Otago). Publication output increased over time; however, study focus appeared to shift from invertebrate to vertebrate species

Table 1. Use of terminology across literature by region

Region	Terminology	No. of uses
Africa	Rehabilitation	7
	Regeneration	1
Asia	Restoration	1
Australasia	Rehabilitation	34
	Restoration	13
	Regeneration	8
	Revegetation	2
	Reclamation	1
Europe	Recultivation	7
	Restoration	3
	Reclamation	3
	Afforestation	2
	Revegetation	1
	Rehabilitation	1
North America	Reclamation	5
	Restoration	2
	Revegetation	1
	Regeneration	1
South America	Rehabilitation	3
	Restoration	2

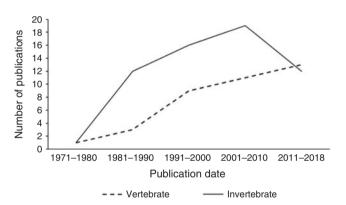


Fig. 1. Publication output for vertebrate and invertebrate responses to mine site restoration studies.

within the last decade (Fig. 1). It is noteworthy that output between any given time bracket is not high within this research area, with a peak rate of less than two papers published annually in the years between 2001 and 2010 (Fig. 1).

Invertebrate responses

Invertebrate responses to mine site restoration were reported in 60 publications (comprising over half (59%) of the literature). Invertebrate studies included species from three phyla (Arthropoda, Annelida, and Mollusca), with a particular focus on the Arthropoda (Insecta; n = 54 of 60). Excluding those assessing multiple groups, studies primarily assessed responses of the Formicidae (ants; n = 19), followed by the Coleoptera (beetles; n = 7), Collembola (springtails; n = 4), Araneae (spiders; n = 3), Diplopoda (millipedes; n = 2), Lepidoptera (butterflies; n = 2), Oligochaeta (earthworms; n = 2), and

Table 2. Summary of target class by mineral type and region for invertebrate studies

Numbers in parentheses denote number of studies for each target or mineral type

Region	Target	Mineral type	
Africa	Multiple invertebrates (2)	Multiple minerals	
	Coleoptera (2)	Sand (1), not stated (1)	
	Diplopoda (2)	Multiple minerals	
Australasia	Formicidae (14)	Bauxite (5), coal (2), sand (4), uranium (2), iron ore (1)	
	Multiple invertebrates (6)	Bauxite (3), coal (1), sand (2)	
	Coleoptera (2)	Peat (1), coal (1)	
	Araneae (2)	Bauxite	
	Collembola (2)	Bauxite	
	Hemiptera (1)	Bauxite	
Europe	Multiple invertebrates (8)	Coal (6), limestone (2)	
•	Coleoptera (3)	Coal	
	Formicidae (2)	Coal (1), not stated (1)	
	Collembola (1)	Coal	
	Oligochaeta (1)	Coal	
North America	Multiple invertebrates (3)	Bentonite (1), peat (1), coal (1)	
	Lepidoptera (2)	Coal	
	Oligochaeta (1)	Bauxite	
South America	Formicidae (3)	Bauxite (1), coal (1), gold (1)	
	Collembola (1)	Sand	

Hemiptera (true bugs; n=1). Twenty studies did not have a focal group and assessed general species diversity and richness for multiple groups. Studies within Australasia and Europe had the widest range of targeted taxa (Table 2). Excluding assessments for multiple invertebrate groups, ants were the most commonly assessed group across almost all mineral types ($\chi^2=49.6$, P<0.001). Of the eight stated mineral operation types (excluding sites listed as 'multiple minerals', or 'not stated'), only three had studies examining more than one invertebrate class (bauxite, coal, and sand mines).

Vertebrate responses

Studies of vertebrate responses to mining restoration comprised less than half of the total number of publications (n = 39of 101, discounting two studies that assessed both invertebrate and vertebrate responses). Studies significantly favoured the use of birds (45%, $\chi^2 = 19.846$, P < 0.001) followed by reptiles (18%, n = 7), mammals (18%, n = 7), and amphibians (3%, n = 7)n=1). Seven studies assessed responses of multiple groups. Of the 39 vertebrate studies only 12 had specific target species, with the other 27 assessing general species diversity and richness. Vertebrate studies primarily originated from Australasia (n = 30), with just three based in each of Europe and North America, and one each in South America, Africa, and Asia. Studies originating outside of Australasia almost exclusively assessed responses of birds, with the exception of three studies (one each in North America, Europe, and Africa) that targeted a combination of mammal, reptile, and amphibian species (Table 3). The type of mineral extracted at sites assessing vertebrate responses to mine site restoration appears to be associated with the region of study. Studies of vertebrate responses at bauxite and sand mines occur exclusively within Australasia, whereas those at coal mines are based either in North America or Europe (Table 3).

Discussion

Studies of faunal responses to mine site restoration are lacking globally, and we found over a 49-year period just 101 peer-reviewed publications reporting on fauna as part of mining restoration activities, with over half from Australia. We interpret this number as 'lacking' because 46 of the 101 studies originated from either the same mining site, or the same locality within a country. Furthermore, as a very rough guide, as of October 2018, Google Scholar reports ~24 000 papers reporting on 'vegetation' AND 'ecological restoration' AND 'mining' in the same period since 1971. Studies of faunal responses to mine site restoration favoured assessments for general species diversity and abundances of invertebrate species. There is a noticeable lack of studies that assess the behaviour and ecology of fauna, particularly of vertebrate species.

Study origin

Australia is at the forefront of mining restoration initiatives, as one of the few countries with widespread legislation (complemented by non-compliance penalties) aimed at mine closure (Gilbert 2000; Clark and Clark 2005; Cristescu et al. 2012). This is reflected in the number of studies reporting faunal responses to mine site restoration originating within Australia. Australia's high activity within the mining restoration field likely results from the increased availability of funding that mineral extraction companies are required to provide for ecological restoration following mine site discontinuation, in order to obtain closure (Clark and Clark 2005). While a leader in restoration research, a recent report identified ~60 000 mine sites across Australia as abandoned (Campbell et al. 2017), of which the number confirmed as restored and officially closed could be as low as 21 (Western Australia: unknown; South Australia: 18 sites; New South Wales, Victoria, and Tasmania: one site each; Queensland and Northern Territory: no confirmed sites:

Table 3. Summary of target taxa by mineral type and continent for vertebrate studies

Numbers in parentheses denote number of studies for each target or mineral type

Region	Class	Target	Mineral type
Mammalia, Amphibia	Aves	Cockatoo (Calyptorhynchus sp.) (2)	Multiple minerals (1), bauxite (1)
		Multiple targets (10)	Bauxite (8), sand (2)
	Mammalia	Swamp wallaby (Wallabia bicolor) (1)	Sand
		Koala (Phascolarctos cinereus) (1)	Sand
		Mouse (Mus sp.) (1)	Sand
		Multiple targets (2)	Sand
		Bat (Chiroptera sp.) (1)	Bauxite
	Reptilia	South-western crevice skink (Egernia napoleonis) (1)	Bauxite
		Bearded dragon (Pogona minor) (2)	Bauxite (1), not stated (1)
		Multiple targets (4)	Bauxite (1), sand (3)
	Amphibia	Multiple targets (1)	Sand
	Mammalia, Reptilia	Multiple targets (2)	Bauxite (1), gold (1)
	Mammalia, Reptilia, Amphibia	Multiple targets (1)	Bauxite
	Reptilia, Amphibia	Multiple targets (1)	Bauxite
North America	Aves	Greater sage-grouse (Centrocercus urophasianus) (1)	Coal
		Multiple targets (1)	Coal
	Amphibia, Reptilia	Multiple targets (1)	Coal
South America	Aves	Multiple targets (1)	Not stated
Europe	Aves	Ring-necked pheasant (<i>Phasianus colchicus</i>), European nightjar (<i>Caprimulgus europaeus</i>), and yellowhammer (<i>Emberiza citronella</i>) (1)	Coal
		Common quail (Coturnix coturnix) (1)	Coal
	Amphibia, Reptilia	Multiple targets (1)	Coal
Asia	Aves	Multiple targets (1)	Tin
Africa	Mammalia	Multiple targets (1)	Not stated

Campbell *et al.* 2017). It is apparent that restoration research focused on reinstatement of fauna after mining is still lacking within Australia. Outside Australia, global mine abandonment numbers are largely either unknown or under-reported. Among countries with (soundly estimated) abandonment figures, high numbers are common, with at least 5000 mine sites in South Africa and 10 000 in Canada identified as abandoned (Cowan *et al.* 2010; Milaras *et al.* 2014), many unlikely to have any substantial ecological management effort that would achieve restoration as defined by McDonald *et al.* (2016).

Rates of mine site cessations and abandonments are cumulatively growing worldwide; however, legislation relating to mine site closure is lacking in most countries (Clark and Clark 2005). Within developed nations, only four countries have widespread legislation relating to mine abandonment (Australia, Japan, Ireland and the United Kingdom), and two have legislation in select states (Canada and the United States: Clark and Clark 2005). Even fewer have legislation for bonding procedures (monetary bond to ensure sites are appropriately restored: Clark and Clark 2005). Just 11 developing countries have complete legislation relating to mine site closure (Clark and Clark 2005), none of which appear in our search results. Globally, Australia appears to be one of the leaders in this space, largely due to comprehensive legislation, although this clearly is not the only motivator as, of the three other developed regions with widespread legislated restoration requirements, we found just one publication relating to faunal responses to mine site restoration (from the United Kingdom). While closure legislation is an essential component in the regulation of mining activities, legislated financial support of restoration activities and research is equally critical.

While much of the literature originates from Australia, almost half of these are from a single organisation: Alcoa's bauxite mining operations in south-west Australia. Not only does this organisation account for a significant proportion of Australian studies, but almost all studies from bauxite mines globally – a mining practice with large surface impacts. These studies originate in a unique ecological region, and a biodiversity hotspot that has been isolated from the rest of the world for a substantial period (Hopper and Gioia 2004). It is highly likely that patterns seen from these studies in the south-west Australian biodiversity hotspot may not provide an accurate representation of faunal responses to mine site restoration in other understudied regions. While it is unlikely that a single, standardised approach to fauna restoration in mining could be implemented globally, due to the ecological diversity of habitats, until legislative requirements and funding increase globally, the diversity of responses by faunal communities to mine site restoration will remain obscure.

Invertebrate responses

Invertebrate species are most commonly studied in assessments of faunal responses to mine site restoration success, and have been studied across a wide diversity of mineral extraction operations. Invertebrates are exceptionally diverse and abundant and typically respond rapidly and with high sensitivity to

habitat disturbance, providing an ideal study group for monitoring environmental change and habitat health (Waltz and Covington 2004; Gerlach *et al.* 2013). Among the mining restoration literature involving studies of particular invertebrate groups, there is a strong focus on assessing diversity and abundances of ant species. Ants have been used extensively as bioindicators in a range of studies, across many habitat types and land uses (Hoffmann and Andersen 2003), including savannahs (Majer 1984b; Andersen 1991; Cross *et al.* 2016a), coastal environments (Majer and Brown 1986; Cross *et al.* 2016b), woodlands and forests (Andersen 1991; Vanderwoude *et al.* 1997), including rainforest (King *et al.* 1998). Ants are an obvious study group of choice, occurring in exceptional abundances in all but three regions (Iceland, Greenland and Antarctica: Hölldobler and Wilson 1990).

Ant community dynamics and responses to disturbances are well studied, and sampling can be performed with ease, rapidity, and at comparatively low cost (Majer 1983; Andersen 1986). One of the few drawbacks in their use stems from difficulties in taxonomy, with many species yet to be described and named (Gerlach et al. 2013). Their widespread use across the mining literature is therefore unsurprising. While ants are the most commonly targeted group, general species diversity assessments for multiple groups (no specific targets) are equally common. General diversity assessments may present further issues, in that they do not account for varying ecologies of species, and identification tends to be broader (Chiarucci et al. 2011). Species diversity and richness assessments are one of the most straightforward and reliable forms of data collection, especially when targeting fauna present in large numbers (Gerlach et al. 2013), likely accounting for the significant bias towards this form of assessment over all other study types.

Vertebrate responses

Vertebrates are less frequently studied in assessments of mine site restoration, and are generally considered to be less effective for use as bioindicators of habitat health than invertebrates (Landres et al. 1988; Bisevac and Majer 1999a; Gerlach et al. 2013). Unlike invertebrates, many species of which occur in high numbers across many habitats, vertebrates can be cryptic, often present in far fewer numbers, and move over greater spatial scales, considerably increasing detection difficulty (Oliver et al. 2009). Few studies assess behavioural and ecological responses of vertebrate fauna, particularly apex predators, to mine site restoration. Behavioural studies can be particularly costly (especially in the initial set-up stage); however, they can also provide extremely successful measures for assessments of the interactions of fauna with their surrounding habitat (Silveira et al. 2003).

Assessments of vertebrate responses to mine site restoration favour avian fauna. This is particularly evident in studies originating outside of Australasia, two-thirds of which assess responses of birds. Birds are relatively easy to detect and identify, have a stabilised taxonomy, often can be common and widespread, and their environmental interactions are well studied, providing an excellent faunal group for use in studies of ecosystem health (Jordano 1982). However, birds may not accurately represent restoration use, as their great mobility may allow for easier recolonisation than other fauna groups.

Second to birds, there are relatively substantial numbers of mammal-focused studies, particularly of charismatic mammals and those that have threatened conservation status. Australia is a land of lizards, and has extremely high rates of endemism (93% endemism: Chapman 2009), yet despite being one of the few countries to assess responses of non-avian taxa, there are surprisingly few reptilian studies. Reptiles are experiencing global declines (Böhm *et al.* 2013), yet they are often overlooked, with few studies examining their response to habitat restoration (Munro *et al.* 2007; Todd *et al.* 2010). Reptiles can provide information on thermal environments (e.g. whether restoration areas have higher associated thermal costs than reference habitats), which other groups, such as birds, may not. Hence, extrapolating responses of birds to poikilothermic fauna is potentially problematic.

Ecosystem function

Research is lacking into ecosystem functionality in terms of assessing interactions of fauna with mine site restoration areas. In many ecosystems, functionality is in some way related to faunal interactions, and loss of biodiversity can greatly impact on ecosystem services (Naeem et al. 1994), yet 81% of studies identified in this review of mine site restoration measure species diversity, abundance, presence, or infer absence. While providing important ecological data, these studies have several drawbacks, and may not provide data on whole ecosystem functionality or be appropriate measures for determining whether a site has been effectively restored. By performing only these assessments, there is a significant chance of missing rare and cryptic species, or in incidental captures of animals moving through the site but not inhabiting the area. This may be particularly problematic in terms of achieving outcomes for mining restoration, as it may provide a false community representation and appear as though a habitat is restored when, in fact, that system may only be in use opportunistically, or not even in use at all.

Moreover, only so much may be learnt from assessing faunal biodiversity. Key ecosystem functions can result or fail as a result of altered animal behaviour and movement patterns (Fahrig 2007; Tarszisz et al. 2018), ecological energetics (Tomlinson et al. 2014), or nutritional physiology (Birnie-Gauvin et al. 2017). This can result in cryptic disruptions to key services such as insect pollination (e.g. Tomlinson et al. 2018) that are not apparent from other studies of pollinator communities such as birds (e.g. Frick et al. 2014). Although there is some evidence that successful mine site restoration is constrained by limited natural recruitment (Koch 2007; James et al. 2011), the role of fauna-mediated pollination and seed dispersal is understudied. Herbivory is a critical plant/animal interaction that has long hampered the restoration of discontinued mining areas, yet has been rarely studied (Keesing and Wratten 1998; Koch et al. 2004; Parsons et al. 2007). These dynamic interactions are important to restoration research, yet fauna are studied only in the context of ecological restoration at a restricted level.

Grey literature and issues with its use

While it is possible that information and data surrounding faunal responses to mine site restoration exist within the grey literature,

we found little empirical data or relevant information within the few that were readily accessible. Accessible grey literature largely comprises premining surveys for fauna species within and around potential new mine sites, conservation and management strategies for rare and threatened species during the life of the mining operation, conference proceedings, or book chapters without empirical data. There is a noticeable dearth of grey literature directly referencing either short- or long-term monitoring of fauna in restoration landscapes, or methods for assessing faunal responses. However, as with published literature, the marginal volume of grey literature to which we could gain access did not discuss fauna in detail, and did not discuss whole animal community return, or return of fully functioning ecosystems. We found eight articles directly referencing fauna in restoration landscapes: three conference proceedings or presentations, three book chapters, and two monitoring plans or guidance documents. Grey literature comprised discussion of the role or return of fauna in mine site restoration (Nawrot and Klimstra 1989; Majer 1997, 1998; Moloney et al. 1998), a monitoring plan for the conservation of rare and threatened fauna (Nickel and Claremont 2015), an assessment of nest translocations for bird species in restoration (termed reclamation) sites (McKee 2007), a guidance document describing techniques for promoting fauna return to rehabilitating sites (Brennan et al. 2005), and a book chapter referencing published studies of vertebrate colonisation of rehabilitating sites at Alcoa (Tibbett 2015). Other resources do recognise the effects of mining on fauna, but this is limited to simple statements on the need for returning habitat components that promote faunal recolonisation; for example, habitat corridors (McLaughlin 2012), monitoring plans for threatened species or management of feral species (without reference to restoration) (Guinea 2007; Weipa 2015; Knuckey 2018) or simply recognition that fauna play important roles in ecosystems and are often overlooked in restoration monitoring (Glenn et al. 2014).

Our biggest challenge in extending our analyses to the grey literature was that resources tend to be largely inaccessible, and often unreliable (Farace and Schöpfel 2010; Corlett 2011). Information and data in unpublished reports and documents are often accessible only within governmental departments and specific regions or countries, and not by the scientific community (Corlett 2011). This has likely resulted in a significant proportion of information within grey literature being overlooked during the development of new conservation and management plans, restoration strategies, and mine site closure policies. It also allows for large, multinational companies to apply different standards in different countries depending on the local legislative and regulatory structures and departments. In order to advance the field of mine site restoration and develop targeted and effective fauna conservation and management strategies, data from these grey literature sources must be peer reviewed, published, and accessible.

Conclusions and future research

The most obvious pattern that has emerged from our review of the literature on responses of fauna to mine site restoration is the overwhelming number of Australian studies contrasted by the surprising dearth of literature for the remainder of the world. This has likely resulted from Australia having both the legislative structure, and financial incentives and capacity for research. To gain an increased understanding of how restoration is impacting ecosystem functioning across a wide range of ecosystems, research must be expanded to a more global level, and encompass a wide range of habitats with varying types of mineral extraction. Not only will this help to account for differences between habitats and ecosystems, but also for the likelihood of varying environmental impact resulting from different mining techniques. Another major limitation is the restricted focus on assessments of behaviour and ecological interactions and functional capacity. Studies of species richness rarely offer insight into the critical ecosystem functions provided by animals. An increased focus must be placed on assessments for ecology and behavioural responses of animals to habitat change and restoration, with an increased emphasis on vertebrate animals within these systems. However, there needs to be a global realisation that mining regulatory systems need to place an emphasis on assessing fauna at multiple taxonomic and functional levels, to ensure that restoration after mining returns an ecosystem to a level of ecological resilience and capacity that matches the local reference ecosystem.

Glossary

Mine discontinuation or abandonment: termination of active mining, ownership of land is retained but site is inactive.

Mine closure: ⁴a whole-of-mine-life process, which typically culminates in tenement relinquishment. It includes decommissioning and restoration' (DMP and EPA 2015).

Conflicts of interest

The authors declare no conflicts of interest.

Acknowledgements

This research was supported by the Australian Research Council Industrial Transformation Training Centre for Mine Site Restoration (ICI150100041). The views expressed herein are those of the authors and are not necessarily those of the Australian Government or the Australian Research Council. The authors thank two anonymous reviewers for their helpful and thorough critique of the manuscript.

References

Altun, A. O., Yilmaz, I., and Yildirim, M. (2010). A short review on the surficial impacts of underground mining. *Scientific Research and Essays* 5, 3206–3212.

Andersen, A. N. (1986). Patterns of ant community organization in mesic southeastern Australia. *Austral Ecology* 11, 87–97. doi:10.1111/J.1442-9993.1986.TB00920.X

Andersen, A. N. (1991). Responses of ground-foraging ant communities to three experimental fire regimes in a savanna forest of tropical Australia. *Biotropica* **23**, 575–585. doi:10.2307/2388395

Andersen, A. N., Hoffmann, B. D., and Somes, J. (2003). Ants as indicators of minesite restoration: community recovery at one of eight rehabilitation sites in central Queensland. *Ecological Management & Restoration* 4, S12–S19. doi:10.1046/J.1442-8903.4.S.2.X

Bian, Z., Dong, J., Lei, S., Leng, H., Mu, S., and Wang, H. (2009). The impact of disposal and treatment of coal mining wastes on environment and farmland. *Environmental Geology* 58, 625–634. doi:10.1007/ S00254-008-1537-0

- Birnie-Gauvin, K., Peiman, K. S., Raubenheimer, D., and Cooke, S. J. (2017).
 Nutritional physiology and ecology of wildlife in a changing world.
 Conservation Physiology 5, 1–18, doi:10.1093/CONPHYS/COX030
- Bisevac, L., and Majer, J. D. (1999a). Comparative study of ant communities of rehabilitated mineral sand mines and heathland, Western Australia. *Restoration Ecology* 7, 117–126. doi:10.1046/J.1526-100X.1999. 72002 X
- Block, W. M., Franklin, A. B., Ward, J. P., Ganey, J. L., and White, G. C. (2001). Design and implementation of monitoring studies to evaluate the success of ecological restoration on wildlife. *Restoration Ecology* 9, 293–303. doi:10.1046/J.1526-100X.2001.009003293.X
- Böhm, M., Collen, B., Baillie, J. E., Bowles, P., Chanson, J., Cox, N., Hammerson, G., Hoffmann, M., Livingstone, S. R., and Ram, M. (2013). The conservation status of the world's reptiles. *Biological Conservation* 157, 372–385. doi:10.1016/J.BIOCON.2012.07.015
- Brennan, K. E. C., Nichols, O. G., and Majer, J. D. (2005). Innovative techniques for promoting fauna return to rehabilitated sites following mining. Australian Centre for Minerals Extension and Research (ACMER), Brisbane, and Minerals and Energy Research Institute of Western Australia (MERIWA Report 248), Perth.
- Bridge, G. (2004). Contested terrain: mining and the environment. Annual Review of Environment and Resources 29, 205–259. doi:10.1146/ ANNUREV.ENERGY.28.011503.163434
- Campbell, R., Linqvist, J., Browne, B., Swann, T., and Grudnoff, M. (2017).Dark side of the boom: what we do and don't know about mines, closures and rehabilitation. The Australia Institute, Canberra.
- Chapman, A. D. (2009). Numbers of living species in Australia and the world. Australian Biological Resources Study, Canberra.
- Chiarucci, A., Bacaro, G., and Scheiner, S. M. (2011). Old and new challenges in using species diversity for assessing biodiversity. *Philo-sophical Transactions of the Royal Society of London. Series B, Biological Sciences* 366, 2426–2437. doi:10.1098/RSTB.2011.0065
- Clark, A. L., and Clark, J. C. (2005). An international overview of legal frameworks for mine closure. Environmental Law Alliance Worldwide: Energy and Biodiversity Initiative.
- Clewell, A., Aronson, J., and Winterhalder, K. (2004). The SER international primer on ecological restoration. SERI (Society for Ecological Restoration International) Science & Policy Working Group.
- Corlett, R. T. (2011). Trouble with grey literature. *Biotropica* 43, 3–5. doi:10.1111/J.1744-7429.2010.00714.X
- Cowan, W., Mackasey, W., and Robertson, J. G. (2010). The policy framework in Canada for mine closure and management of long-term liabilities: a guidance document. Prepared for the National Orphaned/ Abandoned Mines Initiative. Cowan Minerals Ltd., Sudbury, Ontario.
- Cristescu, R. H., Frère, C., and Banks, P. B. (2012). A review of fauna in mine rehabilitation in Australia: current state and future directions. *Biological Conservation* 149, 60–72. doi:10.1016/J.BIOCON.2012.02.003
- Cross, A. T., Myers, C., Mitchell, C. N., Cross, S. L., Jackson, C., Waina, R., Mucina, L., Dixon, K. W., and Andersen, A. N. (2016a). Ant biodiversity and its environmental predictors in the North Kimberley region of Australia's seasonal tropics. *Biodiversity and Conservation* 25, 1727–1759. doi:10.1007/S10531-016-1154-2
- Cross, S. L., Cross, A. T., Merritt, D. J., Dixon, K. W., and Andersen, A. N. (2016b). Biodiversity responses to vegetation structure in a fragmented landscape: ant communities in a peri-urban coastal dune system. *Journal of Insect Conservation* 20, 485–495. doi:10.1007/S10841-016-9881-Y
- Cross, A. T., Young, R., Nevill, P., McDonald, T., Prach, K., Aronson, J., Wardell-Johnson, G. W., and Dixon, K. W. (2018). Appropriate aspirations for effective post-mining restoration and rehabilitation: a response to Kaźmierczak et al. Environmental Earth Sciences 77, 256. doi:10. 1007/S12665-018-7437-Z
- DMP and EPA (2015). Guidelines for preparing mine closure plans.

 Department of Mines and Petroleum, and the Environmental Protection

 Authority, Government of Western Australia, Perth.

- Dudka, S., and Adriano, D. C. (1997). Environmental impacts of metal ore mining and processing: a review. *Journal of Environmental Quality* 26, 590–602. doi:10.2134/JEO1997.00472425002600030003X
- Duffy, J. E. (2003). Biodiversity loss, trophic skew and ecosystem functioning. *Ecology Letters* 6, 680–687. doi:10.1046/J.1461-0248.2003.00494.X
- Ewers, R. M., and Didham, R. K. (2006). Confounding factors in the detection of species responses to habitat fragmentation. *Biological Reviews of the Cambridge Philosophical Society* 81, 117–142. doi:10. 1017/S1464793105006949
- Fahrig, L. (1997). Relative effects of habitat loss and fragmentation on population extinction. *Journal of Wildlife Management* 61, 603–610. doi:10.2307/3802168
- Fahrig, L. (2007). Non-optimal animal movement in human-altered landscapes. Functional Ecology 21, 1003–1015. doi:10.1111/J.1365-2435. 2007.01326.X
- Farace, D., and Schöpfel, J. (2010). 'Grey Literature in Library and Information Studies'. (de Gruyter Saur: Berlin.).
- Frick, K. M., Ritchie, A. L., and Krauss, S. L. (2014). Field of dreams: restitution of pollinator services in restored bird-pollinated plant populations. *Restoration Ecology* 22, 832–840. doi:10.1111/REC.12152
- Gerlach, J., Samways, M., and Pryke, J. (2013). Terrestrial invertebrates as bioindicators: an overview of available taxonomic groups. *Journal of Insect Conservation* 17, 831–850. doi:10.1007/S10841-013-9565-9
- Gilbert, M. (2000). Minesite rehabilitation. Tropical Grasslands 34, 147–154.
- Glenn, V., Doley, D., Unger, C., McCaffrey, N., McKenna, P., Gillespie, M., and Williams, E. (2014). Mined land rehabilitation is there a gap between regulatory guidance and successful relinquishment? AusIMM Bullatin 3, 48
- Gould, S. F., and Mackey, B. G. (2015). Site vegetation characteristics are more important than landscape context in determining bird assemblages in revegetation. *Restoration Ecology* 23, 670–680. doi:10.1111/REC. 12222
- Guinea, M. (2007). Marine turtle management plan for Cape Lambert for Rio Tinto Iron Ore. Version 1.0.
- Hall, H. M., and Grinnell, J. (1919). Life-zone indicators in California. Proceedings of the California Academy of Sciences, 4th series 4, 37–67.
- Hejda, M., Pyšek, P., and Jarošík, V. (2009). Impact of invasive plants on the species richness, diversity and composition of invaded communities. *Journal of Ecology* 97, 393–403. doi:10.1111/J.1365-2745.2009.01480.X
- Hodges, C. A. (1995). Mineral resources, environmental issues, and land use. Science 268, 1305–1312. doi:10.1126/SCIENCE.268.5215.1305
- Hoffmann, B. D., and Andersen, A. N. (2003). Responses of ants to disturbance in Australia, with particular reference to functional groups. *Austral Ecology* 28, 444–464. doi:10.1046/J.1442-9993.2003.01301.X
- Hölldobler, B., and Wilson, E. O. (1990). 'The Ants.' (Harvard University Press: Cambridge, MA.)
- Hopper, S. D. (2009). OCBIL theory: towards an integrated understanding of the evolution, ecology and conservation of biodiversity on old, climatically buffered, infertile landscapes. *Plant and Soil* 322, 49–86. doi:10. 1007/S11104-009-0068-0
- Hopper, S. D., and Gioia, P. (2004). The southwest Australian floristic region: evolution and conservation of a global hot spot of biodiversity. *Annual Review of Ecology Evolution and Systematics* 35, 623–650. doi:10.1146/ANNUREV.ECOLSYS.35.112202.130201
- James, J. J., Svejcar, T. J., and Rinella, M. J. (2011). Demographic processes limiting seedling recruitment in arid grassland restoration. *Journal of Applied Ecology* 48, 961–969. doi:10.1111/J.1365-2664.2011.02009.X
- Jordano, P. (1982). Migrant birds are the main seed dispersers of blackberries in southern Spain. Oikos 38, 183–193. doi:10.2307/3544018
- Kaźmierczak, U., Lorenc, M. W., and Strzałkowski, P. (2017). The analysis of the existing terminology related to a post-mining land use: a proposal for new classification. *Environmental Earth Sciences* 76, 693. doi:10. 1007/S12665-017-6997-7

- Keesing, V., and Wratten, S. D. (1998). Indigenous invertebrate components in ecological restoration in agricultural landscapes. New Zealand Journal of Ecology 22, 99–104.
- King, J. R., Andersen, A. N., and Cutter, A. D. (1998). Ants as bioindicators of habitat disturbance: validation of the functional group model for Australia's humid tropics. *Biodiversity and Conservation* 7, 1627–1638. doi:10.1023/A:1008857214743
- Knuckey, C. (2018). 2017 West Angelas Ghost Bat Monitoring for Rio Tinto Iron Ore. (Biologic Environmental: Perth.)
- Koch, J. M. (2007). Restoring a jarrah forest understorey vegetation after bauxite mining in Western Australia. *Restoration Ecology* 15, S26–S39. doi:10.1111/J.1526-100X.2007.00290.X
- Koch, J. M., Richardson, J., and Lamont, B. B. (2004). Grazing by kangaroos limits the establishment of the grass trees *Xanthorrhoea gracilis* and *X. preissii* in restored bauxite mines in eucalypt forest of southwestern Australia. *Restoration Ecology* 12, 297–305. doi:10.1111/J.1061-2971. 2004.00335.X
- Koch, J. M., Grigg, A. H., Gordon, R. K., and Majer, J. D. (2010). Arthropods in coarse woody debris in jarrah forest and rehabilitated bauxite mines in Western Australia. *Annals of Forest Science* 67, 106. doi:10.1051/ FOREST/2009087
- Lande, R. (1998). Anthropogenic, ecological and genetic factors in extinction and conservation. Researches on Population Ecology 40, 259–269. doi:10.1007/BF02763457
- Landres, P. B., Verner, J., and Thomas, J. W. (1988). Ecological uses of vertebrate indicator species: a critique. *Conservation Biology* 2, 316– 328. doi:10.1111/J.1523-1739.1988.TB00195.X
- Lavelle, P., Decaëns, T., Aubert, M., Barot, S., Blouin, M., Bureau, F., Margerie, P., Mora, P., and Rossi, J. P. (2006). Soil invertebrates and ecosystem services. *European Journal of Soil Biology* 42, S3–S15. doi:10.1016/J.EJSOBI.2006.10.002
- Lin, C., Tong, X., Lu, W., Yan, L., Wu, Y., Nie, C., Chu, C., and Long, J. (2005). Environmental impacts of surface mining on mined lands, affected streams and agricultural lands in the Dabaoshan mine region, southern China. Land Degradation & Development 16, 463–474. doi:10. 1002/LDR.675
- Mace, G. M., Norris, K., and Fitter, A. H. (2012). Biodiversity and ecosystem services: a multilayered relationship. *Trends in Ecology & Evolution* 27, 19–26. doi:10.1016/J.TREE.2011.08.006
- Majer, J. D. (1983). Ants: bio-indicators of minesite rehabilitation, land-use, and land conservation. *Environmental Management* 7, 375–383. doi:10. 1007/BF01866920
- Majer, J. D. (1984b). Recolonisation of ants in rehabilitated open-cut mines in northern Australia. Reclamation and Revegetation Review 2, 279–298.
- Majer, J. D. (1989). 'Animals in Primary Succession: the Role of Fauna in Reclaimed Lands.' (Cambridge University Press: Cambridge.)
- Majer, J. D. (1997). Invertebrates assist the restoration process: an Australian perspective. In 'Restoration Ecology and Sustainable Development'. (eds. K. M. Urbanska, N. R. Webb, P. J. Edwards.) pp. 212–237. (Cambridge University Press: Cambridge.)
- Majer, J. D. (1998). Land reclamation: when will the animals return? In 'Proceedings of the Ecological Society of Australia: Australian Ecosystems: 200 Years of Utilisation, Degradation, and Reconstruction'. (Ed. D. A. Saunders, A. J. M. Hopkins, R. A. How.) pp. 38. (Geraldton, Western Australia).
- Majer, J. D. (2009). Animals in the restoration process—progressing the trends. Restoration Ecology 17, 315–319. doi:10.1111/J.1526-100X. 2009.00528 X
- Majer, J. D., and Brown, K. (1986). The effects of urbanization on the ant fauna of the Swan Coastal Plain near Perth, Western Australia. *Journal* of the Royal Society of Western Australia 69, 13–17.
- Majer, J. D., Orabi, G., and Bisevac, L. (2007). Ants (Hymenoptera: Formicidae) pass the bioindicator scorecard. *Myrmecological News* 10, 69–76.

- McDonald, T., Gann, G., Jonson, J., and Dixon, K. (2016). International standards for the practice of ecological restoration – including principles and key concepts. Society for Ecological Restoration, Washington, DC.
- McKee, G. (2007). Wildlife mitigation techniques at surface coal mines in northeast Wyoming. In 'Proceedings America Society of Mining and Reclamation. National Meeting of the American Society of Mining and Reclamation, Gillette, WY. 30 Years of SMCRA and Beyond, June 2–7, 2007'. (Ed. R. I. Barnhisel.) pp. 425–437. (ASMR: Lexington, KY.)
- McKinney, M. L. (2006). Urbanization as a major cause of biotic homogenization. *Biological Conservation* 127, 247–260. doi:10.1016/J.BIO CON.2005.09.005
- McLaughlin, D. (2012). Rehabilitation strategy. MAC-ENC-MTP-047. BHP Billiton.
- Milaras, M., Ahmed, F., and McKay, T. (2014). Mine closure in South Africa: a survey of current professional thinking and practice. In 'Mine Closure'. (Eds I. M. Weiersbye, A. B. F., M. Tibbett, and K. Mercer.) pp. 1–12. (University of the Witwatersrand: Johannesburg.)
- Miranda, M., Burris, P., Bingcang, J. F., Shearman, P., Briones, J. O., Viña, A., and Menard, S. (2003). 'Mining and Critical Ecosystems: Mapping the Risks.' (World Resources Institute: Washington, DC.)
- Mittermeier, R. A., Turner, W. R., Larsen, F. W., Brooks, T. M., and Gascon, C. (2011). Global biodiversity conservation: the critical role of hotspots. In 'Biodiversity Hotspots: Distribution and Protection of Conservation Priority Areas'. (Eds F. E. Zachos, and J. C. Habel.) pp. 3–22. (Springer: Berlin & Heidelberg.)
- Moloney, D. J., Wilson, B. A., and Kentish, K. (1998). Factors affecting small mammal distribution and abundance in the eastern Otways. II. Fire and mining. In 'Proceedings of the Ecological Society of Australia. Australian Ecosystems: 200 Years of Utilisation, Degradation, and Reconstruction'. (Ed. D.A. Saunders, A.J.M. Hopkins, R.A. How.) pp. 37. (Geraldton, Western Australia.)
- Munro, N. T., Lindenmayer, D. B., and Fischer, J. (2007). Faunal response to revegetation in agricultural areas of Australia: a review. *Ecological Management & Restoration* 8, 199–207. doi:10.1111/J.1442-8903.2007. 00368 X
- Myers, N., Mittermeier, R. A., Mittermeier, C. G., Da Fonseca, G. A., and Kent, J. (2000). Biodiversity hotspots for conservation priorities. *Nature* 403, 853. doi:10.1038/35002501
- Naeem, S., Thompson, L. J., Lawler, S. P., Lawton, J. H., and Woodfin, R. M. (1994). Declining biodiversity can alter the performance of ecosystems. *Nature* 368, 734. doi:10.1038/368734A0
- Nawrot, J. R., and Klimstra, W. D. (1989). Wetland habitat development on mined lands. In 'Animals in Primary Succession: the Role of Fauna in Reclaimed Land'. (Ed. J. D. Majer.) Chapter 11, pp. 269–285. (Cambridge University Press: Cambridge.)
- Nichols, O. G., and Nichols, F. M. (2003). Long-term trends in faunal recolonization after bauxite mining in the jarrah forest of southwestern Australia. *Restoration Ecology* 11, 261–272. doi:10.1046/J.1526-100X. 2003.00190.X
- Nickel, F. A., and Claremont, M. (2015). Conservation significant vertebrate fauna monitoring for Ravensthorpe nickel operations. Terrestrial Ecosystems, Mt Claremont, Western Australia.
- Oliver, P. M., Adams, M., Lee, M. S., Hutchinson, M. N., and Doughty, P. (2009). Cryptic diversity in vertebrates: molecular data double estimates of species diversity in a radiation of Australian lizards (*Diplodactylus*, Gekkota). *Proceedings of the Royal Society of London. Series B, Biological Sciences* 276, 2001–2007. doi:10.1098/RSPB.2008.1881
- Palmer, M. A., Ambrose, R. F., and Poff, N. L. (1997). Ecological theory and community restoration ecology. *Restoration Ecology* 5, 291–300. doi:10.1046/J.1526-100X.1997.00543.X
- Parsons, M. H., Lamont, B. B., Koch, J. M., and Dods, K. (2007). Disentangling competition, herbivory, and seasonal effects on young plants in newly restored communities. *Restoration Ecology* 15, 250–262. doi:10. 1111/J.1526-100X.2007.00208.X

- Perring, M. P., Standish, R. J., Price, J. N., Craig, M. D., Erickson, T. E., Ruthrof, K. X., Whiteley, A. S., Valentine, L. E., and Hobbs, R. J. (2015). Advances in restoration ecology: rising to the challenges of the coming decades. *Ecosphere* 6, 131. doi:10.1890/ES15-00121.1
- R Core Team (2016). R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. Available at: https://www.R-project.org/.
- Ruiz-Jaen, M. C., and Mitchell Aide, T. (2005). Restoration success: how is it being measured? *Restoration Ecology* 13, 569–577. doi:10.1111/J. 1526-100X.2005.00072.X
- Rybicka, E. H. (1996). Impact of mining and metallurgical industries on the environment in Poland. Applied Geochemistry 11, 3–9. doi:10.1016/ 0883-2927(95)00083-6
- Salomons, W. (1995). Environmental impact of metals derived from mining activities: processes, predictions, prevention. *Journal of Geochemical Exploration* 52, 5–23. doi:10.1016/0375-6742(94)00039-E
- Silveira, L., Jacomo, A. T., and Diniz-Filho, J. A. F. (2003). Camera trap, line transect census and track surveys: a comparative evaluation. *Biological Conservation* 114, 351–355. doi:10.1016/S0006-3207(03) 00063-6
- Tarszisz, E., Tomlinson, S., Harrison, M. E., Morrogh-Bernard, H. C., and Munn, A. J. (2018). An ecophysiologically informed model of seed dispersal by orangutans: linking animal movement with gut passage across time and space. *Conservation Physiology* 6, coy013. doi:10.1093/ CONPHYS/COY013
- Tibbett, M. (2015). 'Mining in Ecologically Sensitive Landscapes.' (CSIRO Publishing: Melbourne.)

- Tilman, D., Fargione, J., Wolff, B., D'antonio, C., Dobson, A., Howarth, R., Schindler, D., Schlesinger, W. H., Simberloff, D., and Swackhamer, D. (2001). Forecasting agriculturally driven global environmental change. *Science* 292, 281–284. doi:10.1126/SCIENCE.1057544
- Todd, B. D., Willson, J. D., and Gibbons, J. W. (2010). 'Ecotoxicology of Amphibians and Reptiles.' (CRC Press: Boca Raton, FL.)
- Tomlinson, S., Arnall, S. G., Munn, A., Bradshaw, S. D., Maloney, S. K., Dixon, K. W., and Didham, R. K. (2014). Applications and implications of ecological energetics. *Trends in Ecology & Evolution* 29, 280–290. doi:10.1016/J.TREE.2014.03.003
- Tomlinson, S., Webber, B. L., Bradshaw, S. D., Dixon, K. W., and Renton, M. (2018). Incorporating biophysical ecology into high-resolution restoration targets: insect pollinator habitat suitability models. *Restoration Ecology* 26, 338–347. doi:10.1111/REC.12561
- Vanderwoude, C., Andersen, A. N., and House, A. P. (1997). Ant communities as bio-indicators in relation to fire management of spotted gum (Eucalyptus maculata Hook.) forests in south-east Queensland. Memoirs of the Museum of Victoria 56, 671–675. doi:10.24199/J.MMV.1997.56.69
- Waltz, A. E., and Covington, W. (2004). Ecological restoration treatments increase butterfly richness and abundance: mechanisms of response. *Restoration Ecology* 12, 85–96. doi:10.1111/J.1061-2971.2004.00262.X
- Weipa, R. T. A. (2015). Feral pig management offset strategy. Rio Tinto Iron Ore.
- Wolf, M., and Weissing, F. J. (2012). Animal personalities: consequences for ecology and evolution. *Trends in Ecology & Evolution* 27, 452–461. doi:10.1016/J.TREE.2012.05.001